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| 2 3 4 | Evidence for environmental noise effects on health for the United Kingdom policy context: a systematic review of the effects of environmental noise on mental health, wellbeing, and quality of life; cancer; dementia; birth and reproductive outcomes; and cognition (Supplementary Material) |
|------------------|------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|
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| 10 | |
| 11 | Supplementary Material |
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13 Table S1: Mental health, wellbeing and quality of life extraction table

| Reference | Study Design | Population | Exposure | Comparator | Confounding | Outcome | Findings |
|--------------------------|--------------|----------------------|-----------------------------|-----------------------|--------------------------------|-----------------------|---------------------------------------|
| Weyde, Envt Health, 2017 | Cohort study | Based on the | Road and rail traffic noise | Less than or equal to | Household income, urbanity, | Inattention in 8-year | An association with inattention at |
| | | Norwegian Mother and | exposure was modelled | Lden 30dB | maternal education, ethnicity, | old as reported by | age 8 years was found for road |
| | | Child Cohort Study. | using the Nordic | | maternal alcohol consumption | mothers | traffic noise exposure at age 8 years |
| | | Pregnancy sample: | Prediction Method | | and smoking during pregnancy, | | (coef = .0083, CI = [.0012, .0154]; |

| Feder et al., Environ Res, 2015 | Cross sectional | Randomly selected participants aged 18–79 (606 males, 632 females) | Outdoor wind turbine sound pressure levels were estimated at each dwelling using both ISO9613-1and ISO9613- 2(ISO 1993,1996) as incorporated in the commercial software | 4 dB | Provincial differences | Quality of life, assessed using the WHOQOL-BREF | Wind turbine noise levels were not found to be related to scores on the Physical, Psychological, Social or Environment domains, or to rated QOL and Satisfaction with Health questions |
|--------------------------------------|-----------------|---------------------------------------------------------------------------------------------|----------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|--------|--------------------------------------------------------------------------------------------------------------------------------------------------------------|--------------------------------------------------------------------------|---------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|
| Seidler et al., Environ Res, 2017 | Case control | Individuals aged ≥40 years living in the region of Frankfurt International airport | Cadna Aversion Address-specific exposure to aircraft, road and railway traffic noise in 2005 was estimated | <40 dB | Age, sex, urban living environment and the local proportion of people receiving unemployment benefit as an indicator of socio-economic status | Diagnoses of depression | For road traffic noise, a linear exposure-risk relationship was found for 24-h continuous sound levels ≥70 dB. For aircraft noise, the risk estimates reached a maximum at 50–55 dB and decreased at higher exposure categories. For railway noise, risk estimates peaked at 60–65 dB |
| Welch et al., Noise Health, 2018 | Case control | Residents aged >18 of Wellington city, New Zealand | Residents living within 250 meters of Wellington airport and within 65 Db Ldn contour (Airport Group) | 65 dB | Sex, age, education, employment status, current illness, noise sensitivity | Noise sensitivity – measured using a self- rated three-point scale | People were found to have a significantly poorer Health related QOL than others when they lived near an airport, but not when they lived in the control area |

| Klatte et al., Environ & | Secondary | 1,243 second graders | Aircraft noise levels were | 39 dB | Age, gender, non-verbal | Parent ratings of | Increasing exposure was linearly |
|--------------------------|-----------------|--------------------------|----------------------------|----------------|-------------------------------------|-----------------------|----------------------------------------|
| Behavior, 2016 | analysis of the | from 29 schools around | calculated on the basis of | | abilities, SES, migration | children's quality of | associated with less positive ratings |
| | NORAH dataset | Frankfurt/Main Airport | radar data from the Flight | | background, number of | life and children's | of quality of life, increasing noise |
| | | in Germany | Track and Aircraft Noise | | children's books at home, | wellbeing in school | annoyance, and decreasing reading |
| | | | Monitoring System | | German language proficiency, | | performance. A 20 dB increase in |
| | | | provided by German Air | | percentage of children with a | | aircraft noise exposure was |
| | | | Traffic Services. Road | | migration background in the | | associated with a decrease in |
| | | | traffic and railway noise | | class, mean SES, class size, and | | reading scores of one fifth of a |
| | | | levels were estimated | | parental involvement, classroom | | standard deviation, corresponding |
| | | | using a combination of | | insulation, road-traffic noise, and | | to a reading delay of about 2 |
| | | | information provided by | | railway noise at school | | months |
| | | | local authorities | | | | |
| | | | | | | | |
| Dzhambov et al., Environ | Cross sectional | 720 students aged 18- | Residential noise | 62.4 - 73.5 dB | Sex, age, ethnicity, duration of | Mental health | Evidence that increased residential |
| Res, 2018a | | 35 years, residing in | exposure (LAeq; day | | residence, time spent at | measured using | noise was related to mental ill- |
| | | Plovdiv | equivalent noise level) | | home/day, and stressful life | General Health | health through several indirect |
| | | | was obtained by applying | | events | Questionnaire (GHQ- | pathways |
| | | | a land use regression | | | 12) | |
| | | | (LUR) model | | | | |
| Generaal et al., Psychol | Cross sectional | 2980 participants with | Daily mean noise of road- | N.A | Age, sex, years of education and | The presence of | Neighbourhood socioeconomic |
| Med, 2019 | | and without depressive | rail- and air traffic for | | household income | current diagnoses of | factors, physical factors (high levels |
| | | and anxiety disorders in | several years were | | | depressive disorders | of traffic noise) and social factors |
| | | the past year | modelled by the | | | and anxiety disorders | (lower social cohesion and less |
| | | 1 2 | 5 | | | 5 | ` |
| | | 1 2 | Netherlands | | | | safety) were associated with the |

| | | | Assessment Agency by using the Empara Noise | | | | presence of depressive and anxiety disorders |
|----------------------------------------|----------------------------|------------------------------------------------------------------------------------------------------------------------------------------------|----------------------------------------------------------------------------------------------------------------------------------------------------|-------|-------------------------------------------------------------------------------------------------------------------------------------------------------------|--------------------------------------------------------------------------------------------------------------------------------------|------------------------------------------------------------------------------------------------------------------------------------------------------|
| Dzhambov et al., Environ Res, 2018b | Cross-sectional | 720 students aged 18- 35 years, residing in Plovdiv | Residential noise exposure (LAeq; day equivalent noise level) was obtained by applying a land use regression | N.A | Age, sex, ethnicity, individual level, economic status, duration of residence, time spent at home/day, population, and month of data collection | Mental health measured using General Health Questionnaire (GHQ- 12) | Evidence that having more greenspace near the residence supported mental health through several indirect pathways with serial components |
| Dzhambov et al., Environ Int., 2017 | Cross sectional | 399 students aged 15– 25 years, recruited from two high schools and three universities in Plovdiv | (LUR) model Road traffic noise exposure (Lden) was derived from the strategic noise map of Plovdiv | 50 dB | Sex, age, ethnicity, socioeconomic status and noise sensitivity | Mental health measured using General Health Questionnaire (GHQ- 12) | higher noise exposure was associated with worse mental health only indirectly |
| Zock et al., Environ Int., 2018 | Cross sectional | 4450 registered patients of Dutch GPs who were living in 2013 in one of the 181 five-digit postal code areas in the Netherlands | Exposure to road traffic noise and railway noise was estimated by applying the Standard Model Instrumentation for Noise Assessments | N.A | Sex, age, household income, and socio-economic status and municipality and neighbourhood | Diagnosed (co)morbidity and registered symptoms - coded following the International Classification of Primary Care | A high diversity in land use of neighbourhoods may be beneficial for physical and mental health of the inhabitants |
| Lim et al., Noise Health, 2018 | Population- based study | 918 elementary and middle-school students in South Korea | The level of road traffic noise at the exterior wall of a residential building | N.A | Age, sex, income, premature birth, maternal age at birth, maternal disease during | Mental health in childhood | Noise sensitivity was significantly associated with internalizing, |

| | | | was calculated using noise prediction software | | pregnancy, passive smoking, mental disorders | | externalizing, and total behavioural problems |
|---------------------------------------|-----------------|-------------------------------------------------------------------------------------|----------------------------------------------------------------------------------------------------------|-------|-------------------------------------------------|----------------------|---------------------------------------------------------------------------------------------------------------------------------------|
| | | | based on a noise map | | | | |
| Onchang et al., Noise Health, 2018 | Cross sectional | Student group residing off campus (n= 450) and a control group residing in | Noise levels at both on- campus and off-campus locations were measured using sound level meters | N. A | No adjustments made | GPA score | Various contemporary community noise sources affect university students' activities and possibly influence their educational |
| | | dormitories on-campus | | | | | achievement as well |
| | | (n=336) | | | | | |
| Forns et al., Enviro Health | Cross sectional | Children aged 7-11 | Noise levels inside the | N. A | Child's sex, child's age, black | Total ADHD | Noise was significantly associated |
| Perspectives, 2016 | | years in Barcelona | classroom were measured | | carbon concentrations at home, | symptomatology | with ADHD-DSM-IV scores |
| | | during 2012-13 | during the second 1-week | | traffic noise annovance at home. | | |
| | | uuning 2012 10 | | | | | |
| | | | air pollution sampling | | nome tobacco use, indicators of | | |
| | | | period | | SES at the individual level and | | |
| | | | | | the area level | | |
| He et al., Environ Res., 2019 | Cohort study | 140,456 pregnant | Three indicators of noise | 50 dB | Maternal age, parity, multiple | Hospitalizations for | Compared with 50 dB(A), an LAeq. |
| | | women with no | exposure were used | | pregnancy, stillbirth, | depression or other | 24 h of 60 dB(A) was associated |
| | | documented history of | including A-weighted | | comorbidity, socioeconomic | mental disorders | with 1.16 times (95% CI 0.84–1.62) |
| | | mental illness, who | total outdoor noise | | deprivation, neighbourhood | | the risk of depression |
| | | residing in Montreal | (LAeq. 24 h), day- | | walkability and time period | | hospitalization, and 1.34 times |
| | | | evening-night equivalent | | | | (95% CI 1.04–1.74) the risk of |
| | | | noise (Lden), and | | | | other mental disorders |
| | | | nighttime noise (Lnight) | | | | |

| Civil Aviation Authority, 2017 (Survey of Noise Attitudes) | Cross-sectional study | 1877 adult participants living near Birmingham; East Midlands; Gatwick; Heathrow; London City; Luton; Manchester; Newcastle; | Respondents were selected based on exposure of 51dB LAeq16 hour (92-day average) or higher for summer 2013 using published noise contour | Sampling was stratified so that one- third of the sample was exposed to 51- 54dB LAeq16 hour, and two-thirds were exposed to >54dB | None | Warwick Edinburgh Mental Wellbeing Scale and self- reported health | Found no association between aircraft noise (LAeq 16h 92 day) and wellbeing or self-reported health |
|------------------------------------------------------------------|--------------------------|------------------------------------------------------------------------------------------------------------------------------------------------|------------------------------------------------------------------------------------------------------------------------------------------------------------|------------------------------------------------------------------------------------------------------------------------------------------------------|------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|-----------------------------------------------------------------------------|----------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|
| | | Stansted airports | data for the airport | LAeq16 hour | | | |
| Klompmaker et al., 2019, Environment International | Cross sectional | National public health survey in Netherlands which includes information on 387,195 citizens, aged ≥19 years | Residential traffic noise levels were estimated by the Standard Model Instrumentation for Noise Assessments | N.A | Age, sex, marital status, region of origin, paid occupation, household income, level of education, neighbourhood SES, smoking status, alcohol use, degree of urbanization | Mental health, measured using the Dutch national health survey | Road-traffic noise was only positively associated with prescription of anxiolytics, while rail-traffic noise was only positively associated with psychological distress |
| Enembe et al., 2018, Environmental International | Cross sectional | Eight-thousand Helsinki residents ag 25 years and above were selected from the Population Registry of Finland | Residential exposure to road-traffic noise was estimated from façade noise maps | <45 dB | Sex, age, marital status, employment status, household income, alcohol intake, current smoking status, level of physical activity, pet ownership and sleep disturbance | Use of sleep medication, anxiolytics and antidepressants | Noise annoyance was associated with anxiolytic drug use, OR=1.41 (95% CI: 1.02–1.95), but not with sedative or antidepressant use. There was suggestive association between modelled noise at levels higher than 60 dB and anxiolytic or antidepressant use |

| Oiamo et al., 2015, Social Science and Medicine | Cross sectional | 603 individuals that were exposed to traffic noise and air pollution in Windsor, Ontario, Canada | Residential levels of traffic noise were modelled in SoundPLAN 7.3 | N.A | Sex & age | Health related quality of life measured by the SF-12 health survey | Noise annoyance had a significant and negative effect on both mental and physical health factors of the SF-12 and there was a significant covariance between noise annoyance and odour annoyance |
|--------------------------------------------------------------------------------------------|-----------------|--------------------------------------------------------------------------------------------------------------|-----------------------------------------------------------------------------------------------------------------------------------|-------|-------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|-----------------------------------------------------------------------------|-----------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|
| Leijssen et al., 2019, International Journal of Hygience and Environmental Health | Cross sectional | 23,293 participants, aged 18-70 years, living in Amsterdam between 2011 and 2015 | Modelled daily average noise levels on road- traffic for the year 2011 in the Netherlands using the Empara Noise tool | 45 dB | Age, sex, educational level, occupational status, ethnic origin, marital status, household composition, neuroticism, stressful life events, neighbourhood-level, including socioeconomic status, blue/green space and livability | Depressed mood | Exposure to ≥70 dB(A) compared to the reference group of 45–54 dB(A) showed a significant positive association with depressed mood (OR: 1.65, 95% CI 1.10, 2.48) |
| Zijlema et al., 2015, Int. Journal of Hygiene and Environmental Health | Cross sectional | 5,304 participants, aged between 18-92 years, in the Netherlands | Road traffic noise was estimated using a new implementation of the CNOSSOS-EU noise modelling framework | N. A | Sex, age, educational level and household equivalent income, hostility and vulnerability to stress | Somatic symptoms | No association of noise exposure and somatic symptoms (incidence rate ratio (IRR) 1.001; 95% confidence interval (CI) 1.000– 1.001; n = 56,937) |

| Wallas et al., 2018, Int. Journal of Hygiene and Environmental Health | Cross sectional | 1751 adolescents from the BAMSE birth cohort based in Stockholm County, between 1994-1996 | Traffic noise exposure assessment was performed using data from several national, regional and local authorities | <45 dB | Age, sex, rhinitis, eczema and sampling season | Saliva cortisol levels | Road traffic noise exposure was not associated with saliva cortisol, however, annoyance to noise tended to increase the levels. Saliva cortisol levels appeared particularly high among those who were highly annoyed and exposed to road traffic noise levels≥55 dB Lden |
|-----------------------------------------------------------------------------|-----------------|-------------------------------------------------------------------------------------------------------------------------------------------------------|----------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|--------|--------------------------------------------------------------------------------------------------------------------------------------------------------------------|------------------------------------------------------|------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|
| Lawton et al., 2016, Transportation Research Part D | Cross sectional | Two-year sample of nearly 190,000 households, from the Annual Population Survey, UK | Noise contour data were derived from annual average noise levels from 2012 airport operator strategic noise maps at the geographical level of residential dwelling outer area | 55 dB | Ethnicity, household income, health status, marital status, employment status, housing status, gender, age, geographic region, religion, and education | Subjective wellbeing | The presence of daytime aviation noise was found to consistently negatively impact on five subjective wellbeing measures |
| Wright et al. 2018, Environmental Health | Cross sectional | 198,532 people enumerated at the 2011 Census, aged 18 years and over and living within the 54 dB Belfast City Airport noise contour | Residential exposure to aircraft noise (LAeq,16h) was assessed by linking Census records with modelled noise contours surrounding George Best Belfast City Airport | 54 dB | Age, sex, ethnicity, religion and marital status, education, property value and car availability, likelihood of poor mental | Prevalence of self- assessed mental ill health | No association between aircraft noise and risk of mental ill health |

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|----|-------------------------|--------------|-------------------------|----------------------------|-----|---------------------------------|------------------------|--------------------------------------|
| v | an Aart et al., Environ | Longitudinal | 172 Belgian children | A GIS-based noise model | N.A | Age, sex & socioeconomic status | Childhood | Inverse association between |
| In | nt., 2018 | | aged 6.7-12.2, followed | including the Flemish | | | psychosocial stress - | residential and traffic density with |
| | | | for three years (2012- | street and railway | | | Strengths and | hyperactivity problems |
| | | | 2015) | networks was used to | | | Difficulties | |
| | | | | estimate traffic noise | | | questionnaire and hair | |
| | | | | levels in 5 dB(A)- | | | cortisol | |
| | | | | intervals according to the | | | | |
| | | | | European Noise Directive | | | | |
| | | | | (2002/49/EC) | | | | |
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3031 Table S2: Cancer extraction table

| Reference | Study Design | Population | Exposure | Comparator | Confounding | Outcome | Findings |
|--------------------------------------------------------------|------------------------------|--------------------------------------------------------------------------------------------------------------------------------------------|------------------------------------------------------------------------------------------------------------------------------------------------------------------|------------|-----------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|-------------------------------------------|----------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|
| Andersen et al., Lynge Breast Cancer Res, 2018 | Longitudinal cohort study | 22,466 female nurses (age > 44 years) who at recruitment in 1993 or 1999 reported information on breast cancer risk factors | Road traffic noise levels at the nurse road traffic noise levels at the nurses' residential addresses were calculated using the Nord2000 | <48 dB | Age, birth cohort, body mass index, alcohol use, leisure time physical activity, smoking status, age at menarche, parity (yes; no), number of children, age at first birth, menopausal status, HT use, and oral contraceptive use | Incidence of breast cancer | For each 10 dB increase in 24-year mean noise levels at their residence, a statistically significant 10% increase in total breast cancer incidence was found |
| Hegewald et al., Scandinavian J Work Envt Health, 2017 | Prospective | 1,026 ,670 Women aged ≥40 years by 2010 living in the region surrounding the Frankfurt international airport | Aircraft noises, road traffic noise and rail noise were all modelled using SoundPLan. | <40 dB | Age, hormone replacement therapy, and regional proportion of people receiving unemployment benefits | Incident diagnoses of breast cancer | An increased odds ratio was observed for estrogen receptor negative tumors at 24- hour aircraft noise levels 55–59 dB [OR 55–59 dB 1.41. Clear associations between road and rail traffic noise were not observed |

| Roswall et al., Environ | Longitudinal | 57,053 participants | Residential road | N. A | Calendar year at diagnosis, train noise, | Overall | No association was found between time- |
|--------------------------|--------------|------------------------|-----------------------|--------|------------------------------------------|-------------------|----------------------------------------------|
| Reseasrch, 2016 | cohort study | (29,875 women) aged | traffic noise was | | smoking, alcohol intake, abstainers, | mortality and | weighted averages of residential road |
| | | 50-64 years of age | calculated as the | | waist circumference, recreational | breast cancer- | traffic noise 1-,3- or 5- years before death |
| | | residing in Copenhagen | equivalent continuous | | physical activity and marital status | specific | and overall or breast cancer-specific |
| | | | A-weights sound | | | mortality | mortality |
| | | | pressure level (LAeq) | | | | |
| | | | | | | | |
| Roswall, et al., Cancer, | Longitudinal | 57,053 participants | Traffic noise was | <40 dB | Age, sex, railway noise, smoking, | Overall | No association found between residential |
| Causes & Control, 2017 | cohort study | (29,875 women) aged | calculated for all | | smoking duration, smoking intensity, | colorectal cancer | road traffic noise and rectal cancer. |
| | | 50-64 years of age | residential addresses | | alcohol intake, abstainers. recreational | incidence | Observed an association with distal colon |
| | | residing in Copenhagen | from 1987 to 2012 for | | physical activity, education, whole | | cancer, but not for proximal colon cancer: |
| | | | 51,283 Danes in the | | grain cereal, red meat, and marital | | 0.99 (0.83-1.18), per 10 dB, 10 years |
| | | | Diet, Cancer and | | status, income and municipal-level | | preceding diagnosis. There was no |
| | | | Health Cohort. | | population density at baseline | | association between railway noise and |
| | | | Railway noise also | | | | colorectal cancer, or any subtype |
| | | | calculated | | | | |
| | | | | | | | |

| Roswall et al., PloS One, 2015 | Longitudinal cohort study | 27,178 men aged 50-65 years born in Denmark with no previous cancer diagnosis and living in Greater Copenhagen | Road and railway traffic was calculated using SoundPLAN | N. A | Education level, area level socioeconomic position of baseline municipalities or districts for Copenhagen municipality based on municipality/district information on | Incidence rate ratios for association between road traffic and | There was no association between residential road traffic noise and risk of prostate cancer for any of the three exposure windows. For railway noise, there was no association with overall |
|-----------------------------------|------------------------------|----------------------------------------------------------------------------------------------------------------------------|---------------------------------------------------------------|------|----------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|----------------------------------------------------------------------------|--------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|
| | | | | | education; work market affiliation; income; smoking status; smoking duration; body mass index; waist circumference; physical activity; calendar year; and airport noise | railway noise and prostate cancer | prostate cancer |
| Roswall et al., PloS One, 2017 | Longitudinal cohort study | 57,053 participants (29,875 women) aged 50-64 years of age residing in Copenhagen | Road and railway traffic was calculated using SoundPLAN | N. A | Age, calendar year of diagnosis and sex, railway noise at diagnosis, baseline smoking status, baseline smoking duration, baseline alcohol intake, baseline abstainers, baseline red meat intake, baseline recreational physical activity, education 1 year before diagnosis and income 1 year before diagnosis | Overall mortality and colorectal cancer-specific mortality | No association was found between road traffic noise and overall (MRR 1.00 (0.88-1.13) per 10dB) or colorectal cancer – specific mortality (MRR 0.98 (0.85-1.13) per 10 dB) over the entire follow-up period, or 1 years preceding death. Railway noise was only included as a covariate |

| Sorensen et al., I J of | Longitudinal | (29,875 women) aged | Road traffic and | <42 dB | Age, parity, age at first birth, hormone | Incidence rate | No overall association was found |
|-------------------------|--------------|------------------------|-----------------------|--------|------------------------------------------|-------------------|--------------------------------------------|
| Cancer, 2014 | cohort study | 50-64 years of age | railway traffic noise | | replacement therapy status and | ratios (IRRs) for | between residential road traffic or |
| | | residing in Copenhagen | exposure were | | duration, age at menarche, length of | breast cancer in | railway noise and breast cancer risk. |
| | | | calculated using | | school attendance, BMI, alcohol | association with | Among women with estrogen receptor |
| | | | SoundPLAN | | consumption, alcohol intake, smoking | road traffic and | negative breast cancer, a 10-dB higher |
| | | | | | status, intake of vegetables, physical | railway noise | level of road traffic noise during the |
| | | | | | activity, calendar-year and railways | | previous 1, 5 and 10 years were |
| | | | | | and airport noise | | associated with 28%, 23% and 20% |
| | | | | | | | higher risks of estrogen receptor negative |
| | | | | | | | breast cancer. Similarly, a 10-dB increase |
| | | | | | | | in railway noise increased risk for |
| | | | | | | | estrogen receptor negative breast cancer |
| | | | | | | | by 38%. No association was found |
| | | | | | | | between road traffic or railway noise and |
| | | | | | | | estrogen receptor positive breast cancer |

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| Sorensen et al., | Case control | 2753 Cases were | Road traffic noise | 55 dB | Age and sex, education, disposable | Odds ratios and | A 5—year time-weighted mean of road |
|-------------------------|--------------|-------------------------|------------------------|-------|---------------------------------------|--------------------|-------------------------------------------|
| Environmental Research, | | identified using the | exposure was | | income, cohabitation status, Charlson | 95% confidence | traffic noise about 65dB was associated |
| 2015 | | Cancer Registry. | calculated at the most | | comorbidity index, air pollution | internals for risk | with an 18% higher risk for non-hodgkin |
| | | Eligible cases were | exposed façade for all | | | for non-hodgkin | lymphoma (NHL) when compared to |
| | | Danes between 30 and | present and historical | | | lymphoma | road traffic noise below 55dB, whereas |
| | | 84 years of age with a | addresses using | | | associated with | for exposure between 55 and 65dB no |
| | | primary diagnosis of | Sound PLAN | | | exposure to | association was found. In analyses of |
| | | NHL between 1992 and | | | | traffic noise | NHL subtypes, no association was found |
| | | 2010. For each case, | | | | | between road traffic noise and risk of T- |
| | | two random controls, | | | | | cell lymphoma, whereas increased risks |
| | | matched on sex and | | | | | for B-cell lymphoma and unspecified |
| | | year of birth were | | | | | lymphomas were observed as exposures |
| | | selected from the Civil | | | | | above 65dB |
| | | Registration System | | | | | |
| | | | | | | | |

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43 Table S3: Dementia and other neurodegenerative outcomes extraction table

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| Reference | Study Design | Population | Exposure | Comparator | Confounding | Outcome | Findings |
|--------------------------------|------------------|-------------------------|-----------------------------|----------------|--------------------------------------|-----------------------------|--------------------------------|
| Andersson et al., | Longitudinal | 1721 participants from | Modelled data provided | 35 dB | Age, ApoE4, education, physical | Dementia incidence | Exposure to noise |
| Environmental Research, | cohort study | the Betula project, 985 | road traffic noise levels | | activity, smoking, sex, alcohol use, | | levels (Leq. 24 h)> 55 dB had |
| 2018 | | men and 736 women, | (Leq. 24 h) at the | | BMI, and waist-hip ratio, | | no significant effect on |
| | | who at baseline were | participants' residential | | hypertension, diabetes, and stroke | | dementia risk (HR 0.95; CI: |
| | | aged 55-85 years | address at baseline | | | | 0.57, 1.57). |
| | | (mean 68.5 ± 9.4) | | | | | |
| Constrated DMLOner 2018 | P atmost a ativa | 120.078 adults aged | Traffic intersity distance | Maan avn aguna | Ass say otherisity smalling and | A first recorded discussion | There was a monitive evenesure |
| Carey et al., Bivij Open, 2018 | Retrospective | 150 978 aduns aged | Traffic Intensity, distance | Mean exposure | Age, sex, ennicity, smoking and | A first recorded diagnosis | There was a positive exposure |
| | cohort study | 50-79 years registered | from major road and | Lnight 52.1dB | body mass index, area deprivation | of dementia and, where | response relationship between |
| | | with their general | night-time noise levels | | and comorbidity (IHD, stroke, | specified, subgroups of | dementia and all measures of |
| | | practices on 1 January | (Lnight) were estimated | | diabetes, heart failure) | Alzheimer's disease and | air pollution except O3. |
| | | 2005, with no | at the postcode level | | | vascular dementia during | Increases in dementia risk |
| | | recorded history of | linked to clinical data via | | | 2005–2013 | were also observed with |
| | | dementia or care home | residential postcode. | | | | PM2.5, PM2.5 specifically |
| | | residence | Road traffic noise levels | | | | from primary traffic sources |
| | | | were estimated using the | | | | only and Lnight, but only NO2 |
| | | | TRAffic Noise EXposure | | | | and PM2.5 |
| | | | (TRANEX) model | | | | |
| | | | | | | | |

| Culqui et al., Science of Total Environment, 2017 | Longitudinal ecological time series study | Madrid during the period 2001-2009 – mean population of 3,116,897 and of this total, 754,005 persons (24.2%) were aged 60 years or over | Mean daily noise levels (dB(A)) for equivalent diurnal noise level 7–23 h (Leqd), equivalent nocturnal noise level 23– 7 h (Leqn), and daily noise level 24 h (Leq24) were measured | N. A | No adjustments made | Short-term admissions to hospital for Alzheimer's disease (ICD-9 code) | There was no statistically significant association with emergency Alzheimer's disease admissions or noise |
|------------------------------------------------------|---------------------------------------------------|-------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|-------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|------|---------------------|--------------------------------------------------------------------------------------------------|------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|
| Linares et al., Environ Res.,2017 | Longitudinl ecological time-series study | Number of daily dementia-related emergency (DDE) hospital admissions to Madrid municipal as obtained from the Hospital Morbidity Survey (National Statistics Institute) | Leqd, equivalent diurnal noise level (from 8 to 22 h), and Leqn, equivalent nocturnal noise level (from 22 to 8 h) in dB(A) was provided by the Madrid Municipal Air Quality Monitoring Grid | N. A | | Number of daily dementia- related emergency admissions to municipal hospitals in Madrid | Admissions displayed a linear functional relationship without a threshold in the case of Leqd. The RR of DDE admissions was 1.15 (1.11–1.20) for an increase of 1 dB in Leqd |

| Diaz J et al., 2017 Gac Sanit | Ecological time | The population of | The Madrid Municipal | 55 dB | Temperature, pollution, trends and | Parkinson's Disease related | The association between Leqd |
|-------------------------------|-----------------|--------------------------|----------------------------|-------|-------------------------------------|-----------------------------|---------------------------------|
| | series analysis | Madrid during the | Air Quality Monitoring | | seasons | demand for healthcare | and Hospital admissions was |
| | | period 2001-2009, it | Grid supplied Leqd, | | | | found to be linear. Leqd and |
| | | had a mean population | equivalent diurnal noise | | | | Leqn at lag 0.1 and |
| | | of 3,116,897 and of this | level (from 8 to 22 h), | | | | temperature at lags 1 and 5 |
| | | total, 284,929 persons | and Leqn, equivalent | | | | were the only environmental |
| | | (9%) were aged 75 | nocturnal noise level | | | | variables associated with |
| | | years or over | | | | | increased Parkinson's disease |
| | | | | | | | related healthcare demand |
| | | | | | | | |
| Carmona et al., 2018 Science | Longitudinal | The population of | Mean daily noise levels | N. A | Linear trends, seasonality and the | Number of emergency MS | Traffic noise can exacerbate |
| of the Total Environment | ecological time | Madrid during the | (dB(A)) for equivalent | | autoregressive nature of the series | hospital admissions | MS symptoms, leading to |
| | series study | period 2001-2009, it | diurnal noise level 7–23 h | | itself | | hospital admissions due to this |
| | | had a mean population | (Leqd), equivalent | | | | cause |
| | | of 3,116,897 | nocturnal noise level 23- | | | | |
| | | | 7 h (Leqn), and daily | | | | |
| | | | noise level 24 h (Leq24), | | | | |
| | | | supplied by the Madrid | | | | |
| | | | Municipal Air Quality | | | | |
| | | | Monitoring Grid | | | | |
| | | | | | | | |
| | | | | | | | |

| | Tzivian et al., Environmental Health Perspectives, 2016 | Cross-sectional cohort study | 4,086 participants who were 50–80 years old | Lden & Lnight – long term exposure at the baseline address (2000- 2003) | Lden mean 53.74 (SD+- 9.49) Lnight mean 44.88 (SD+- 9.17) | Age, sex, socioeconomic status, alcohol consumption, smoking status, self-reported environmental tobacco smoke, any regular physical activity, body mass index | Diagnoses of Mild cognitive impairment (MCI) | A 10 dB(A) increase in LDEN was associated with overall MCI. A 10dB(A) increase in Lnight was associated with overall MCI |
|----------|------------------------------------------------------------|---------------------------------|------------------------------------------------|----------------------------------------------------------------------------------|--------------------------------------------------------------------------|----------------------------------------------------------------------------------------------------------------------------------------------------------------------------|----------------------------------------------------|---------------------------------------------------------------------------------------------------------------------------------------|
| 45 46 | 5 | | | | | | | |
| 47 | 7 | | | | | | | |
| 48 | 3 | | | | | | | |
| 49 |) | | | | | | | |
| 50 |) | | | | | | | |
| 51 | | | | | | | | |
| 52 | | | | | | | | |
| 53 | 3 | | | | | | | |
| 54 | ŀ | | | | | | | |
| 55 | 5 | | | | | | | |
| 56 |) | | | | | | | |
| 57 | 7 | | | | | | | |
| 58 | | | | | | | | |
| 59 |) | | | | | | | |
| 60 |) | | | | | | | |

61 Table S4: Birth and reproductive outcomes extraction table

| Reference | Study Design | Population | Exposure | Comparator | Confounding | Outcome | Findings |
|-----------------------------|--------------|--------------------------|-------------------------|------------|---------------------------------------------|--------------------------|---------------------------------|
| | | Study based on children | Estimated annual levels | N. A | Age, sex of child, maternal age at birth, | Children with febrile | An interquartile range increase |
| Hjortebjerg et al., Scand J | Cohort study | enrolled in the Danish | of road traffic and | | parity, smoking during first trimester, | seizures were identified | in childhood exposure to road |
| Work Environ Health, 2018 | | National Birth Cohort | railway noise at all | | alcohol consumption during first | by linking the unique | traffic noise and air pollution |
| | | which consecutively | addresses using | | trimester, level of education, disposable | personal identification | was associated with an 11% |
| | | recruited pregnant women | SoundPLAN | | income one year before birth of the | number of each child in | and 5% higher risk for febrile |
| | | from March 1996 to | | | child, road traffic noise for air pollution | the study base to the | seizures, respectively, after |
| | | November 2002 from all | | | and vice versa | nationwide Danish | adjustment for potential |
| | | over Denmark. 57,282 | | | | National Patient | confounders |
| | | children participated | | | | Register | |
| | | | | | | | |
| | | | | | | | |
| | | | | | | | |
| | | | | | | | |

| Min KB & Min JY, Environ | Population | Used the National Health | Data on noise exposure | 35 dB | Age, income, residence area, smoking | Diagnoses of male | A non-linear dose-response |
|--------------------------|--------------|----------------------------|------------------------|-------|------------------------------------------|-------------------------|---------------------------------|
| Pollut, 2017 | based cohort | Insurance Service- | was obtained from the | | history, exercise, alcohol use, blood | infertility | relationship was observed |
| | study | National Sample Cohort | National Noise | | glucose levels, BMI, history of diseases | | between infertility and |
| | | (2002-2013), a | Information System | | | | quartiles of daytime and night |
| | | population-wide health | | | | | time noise after adjustment for |
| | | insurance claims dataset. | | | | | confounding variables. Based |
| | | A total of 206,492 males | | | | | on WHO criteria, adjusted |
| | | of reproductive age (20-59 | | | | | odds for infertility were |
| | | years) with no history of | | | | | significantly increased (OR 1/4 |
| | | congenital malformations | | | | | 1.14; 95% CI, 1.05e1.23) in |
| | | | | | | | males exposed to night time |
| Pedersen et al., Environ | Cohort study | 84,218 liveborn singletons | Road traffic noise was | N. A | Maternal smoking, maternal alcohol | Diagnoses of congenital | Residential road traffic |
| Res., 2017 | | (1997–2002) from the | calculated at the most | | consumption, parental age (years), | anomalies | exposure to noise or air |
| | | Danish National Birth | exposed facade of each | | maternal education the year before last | | pollution during pregnancy did |
| | | Cohort | residential address | | menstrual period, household disposable | | not increase |
| | | | using SoundPLAN | | income, parity, maternal pre-pregnancy | | risk for development of |
| | | | | | BMI | | congenital anomalies |
| | | | | | | | |
| | | | | | | | |
| | | | | | | | |

| | Retrospective | 540,365 singleton term | A-weighted road traffic | 50 dB | Maternal age, birth registration type, | Birth weight outcomes | Trends of decreasing birth |
|-------------------------|---------------|-----------------------------|-------------------------|-------|-----------------------------------------|---------------------------|---------------------------------|
| Smith et al., BMJ, 2017 | population- | live births | noise levels (dB) were | | birth season, birth year, Carstairs | - low birth weight, | weight across increasing |
| | based cohort | | modelled to 0.1 dB | | deprivation quintile, tobacco | small for gestational age | road traffic noise categories |
| | study | | resolution for all | | expenditure. Birth weight and LBW | | were observed, but were |
| | | | geocoded maternal | | were adjusted for sex, gestational age | | strongly attenuated when |
| | | | residential addresses | | and baby's ethnicity | | adjusted for primary traffic |
| | | | using the Traffic Noise | | | | related air pollutants |
| | | | Exposure (TRANEX) | | | | |
| | | | model | | | | |
| Dzhambov et al., 2019, | Explorative | Used data from two cross- | Noise emissions were | 10 dB | sex of child, age of mother at birth, | Birth outcomes – low | An increase of Ldn was |
| Science of the Total | study | sectional studies (UIT, n = | calculated in | | gestational age, single mother status, | birth weight and small | associated with higher odds for |
| Environment | | 573 and BBT, n = 518) in | 2003/2004.Total day- | | mother's education, smoking during | for gestational age | low birth weight but only in |
| | | the Tyrol Region | evening night noise | | pregnancy, duration of residence before | | one of the studies. |
| | | (Austria/Italy) | level (Lden) was | | conception, and house type, Ldn or | | Unexpectedly, an increase in |
| | | | calculated at the most | | NO2 | | Lden was associated with an |
| | | | exposed façade | | | | increase in the odds for being |
| | | | covering road and rail | | | | small for gestational age. |
| | | | noise | | | | |

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| Poulsen et al., 2018, | Nationwide | 135,795 pregnant women | Estimated hourly | >24 dB | Sex, calendar year of birth, maternal | Small for gestational | No association between night- |
|------------------------------|--------------|----------------------------|---------------------------|--------|------------------------------------------|--------------------------|-------------------------------------|
| Environmental Research | registered | living in Danish dwellings | outdoor and low | | age at birth, parity, season of | age, pre-term birth and | time wind turbine noise and |
| | based study | from 1982 to 2013 | frequency indoor wind | | conception, marital status, education, | birth weight | adverse birth outcomes |
| | | | turbine noise at the | | work status, personal income, area-level | | |
| | | | dwellings of the | | mean disposable income, ever living on | | |
| | | | pregnant women and | | a farm during pregnancy, shortest | | |
| | | | aggregated as mean | | distance to road with ≥5000 vehicles | | |
| | | | night-time Wind | | per day during pregnancy, traffic load | | |
| | | | turbine noise during | | within 500m radius of dwelling | | |
| | | | pregnancy | | averaged over all addresses held during | | |
| | | | | | pregnancy, maternal smoking during | | |
| | | | | | 1st trimester and for both maternal | | |
| Wallas et al., Environ Res., | Longitudinal | 4089 children born 1994- | Road traffic noise | N. A | Parental occupation, smoking during | Data on BMI from birth | Residential road traffic noise |
| 2019 | cohort study | 1996 in four pre-defined | levels were estimated | | pregnancy, maternal BMI and | to adolescence were | exposure was associated with |
| | | areas of Stockholm | at the most exposed | | municipality at birth | collected via | increases in BMI from school age |
| | | County | façade of all residential | | | questionnaires, clinical | to adolescence, but not at earlier |
| | | | homes where the study | | | examinations and health | ages. Maternal noise exposure |
| | | | subjects lived. | | | care records. A national | during pregnancy was generally |
| | | | Maternal occupational | | | register provided | unrelated to adverse birth outcomes |
| | | | noise exposure during | | | information on birth | (low birth weight, pre-term birth) |
| | | | pregnancy was | | | outcomes | and BMI from birth to adolescence |
| | | | estimated based on | | | | in the children, however, traffic |
| | | | self-reported | | | | noise exposure was associated with |
| | | | occupation | | | | a decreased risk of preterm birth |

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65 Table S5: Cognition extraction table

66

| Reference | Study Design | Population | Exposure | Comparator | Confounding | Outcome | Findings |
|----------------------------|-----------------|--------------------------|-----------------------------|------------|---------------------|---------------------|-----------------------------------------|
| | | | | | | | |
| | | | | | | | |
| Papanikolaou et al., Int J | Cross sectional | 676 participants (324 | Selected schools on the | 55-66 dB | No adjustments made | To assess the | Children in low-level noise schools |
| Adolesc Med Health, 2015 | | boys,47.9% and 352 | basis of increasing levels | | | effects of noise on | showed statistically significant |
| | | girls, 52.1%) of the 4th | of exposure to road traffic | | | cognitive skills, a | differences from children in medium- |
| | | and 5th elementary | noise and classified into | | | test was | and high-level noise schools in reading |
| | | classes | three categories, | | | constructed | performance (p < 0.001). Children in |
| | | | according to external | | | based on the | low-level noise schools differed |
| | | | noise: Low-level noise: | | | National | significantly from children in high- |
| | | | 55–66 dB, Medium-level | | | Curriculum for | level noise schools but only in |
| | | | noise: 67–77 dB, and | | | Elementary | mathematics performance ($p = 0.001$) |
| | | | High-level noise: 72-80 | | | Education for | |
| | | | dB. Noise levels outside | | | reading and | |
| | | | classrooms were | | | mathematics | |
| | | | measured with an echo | | | | |

| Tzivian et al., Environ Health Perspectives 2016 | Cross sectional | 4,814 randomly chosen men and women who were 45–75 years old at baseline, enrolled into the study between December 2000 and August 2003 | Long-term exposure to traffic noise was modeled according to the European Directive 2002/49/EC as the weighted 24-hr mean (LDEN) and the night- time mean (LNIGHT) at the baseline address | N. A | Age, sex, socioeconomic status, alcohol consumption, smoking status, self-reported environmental tobacco smoke, any regular physical activity, body mass index, background NO2 | Assessment of overall mild cognitive impairment (MCI) and amnestic (aMCI) and nonamnestic (naMCI) mild cognitive | Most air pollutants and traffic noise were associated with overall MCI and aMCI. 10 A-weighted decibel [dB(A)] increase in LDEN was associated with overall MCI |
|-----------------------------------------------------|-----------------|-----------------------------------------------------------------------------------------------------------------------------------------------------------|--------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|------|-----------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|------------------------------------------------------------------------------------------------------------------------------------------|-------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------------|
| Tzivian et al., J Toxicol Environ Health A, 2017 | Cross sectional | 4814 randomly chosen men and women aged between 45 and 75 years | Long-term exposure to traffic noise was modeled according to the European Directive 2002/49/EC | N. A | Age, sex, socioeconomic status, alcohol consumption, smoking status, ETS, any regular physical activity, and BMI | impairment Cognitive function – measured through a cognitive performance assessment | High noise exposure increased the association of air pollution with cognitive function. observed stronger negative associations in participants with double exposure compared to the addition of effect estimates of each single exposure |

| Klatte et al Environ & | Secondary | 1 243 second graders | Aircraft noise levels were | 30 dB | Age gender non-verbal abilities | Reading | Increasing exposure was linearly |
|-----------------------------|-----------------|--------------------------|-----------------------------|-------|-----------------------------------------|-------------------|-------------------------------------------|
| | | | | 57 dD | | C | |
| Benavior, 2016 | analysis of the | from 29 schools around | calculated on the basis of | | SES, migration background, | performance - | associated with less positive ratings of |
| | NORAH dataset | Frankfurt/Main Airport | radar data from the Flight | | number of children's books at | assessed though a | quality of life, increasing noise |
| | | in Germany | Track and Aircraft Noise | | home, German language | standardised | annoyance, and decreasing reading |
| | | | Monitoring System | | proficiency, percentage of | reading | performance. A 20 dB increase in |
| | | | provided by German Air | | children with a migration | comprehension | aircraft noise exposure was associated |
| | | | Traffic Services. Road | | background in the class, mean | test for primary | with a decrease in reading scores of one |
| | | | traffic and railway noise | | SES, class size, and parental | school children | fifth of a standard deviation, |
| | | | levels were estimated | | involvement, classroom | instructed in | corresponding to a reading delay of |
| | | | using a combination of | | insulation, road-traffic noise, and | German | about 2 months |
| | | | information provided by | | railway noise at school | | |
| | | | local authorities and were | | | | |
| | | | used only as covariates | | | | |
| | | | | | | | |
| Spilski et al., ICBEN, 2017 | Cross sectional | 1,243 | Exposure levels at schools | 34 dB | Age, gender, SES, migration | Cognition and | A 10 dB increase of aircraft noise at |
| | | children participated in | and at the children's homes | | background, German language | quality of life | school was associated with a decrease |
| | | the study | were assessed by the | | proficiency, number of children's | | in children's global reading and word |
| | | | NORAH acoustic team. | | books, non-verbal abilities, story | | reading scores by one tenth of a SD - |
| | | | Aircraft noise levels were | | comprehension, phonological | | one point on the T-score scale. For text |
| | | | calculated on the basis of | | awareness, access to phonological | | reading, a 10 dB increase of aircraft |
| | | | radar data from the Flight | | representations, class socioeconomic | | noise was associated with a decrease by |
| | | | Track and Aircraft Noise | | status (SES), class size, percentage of | | one eighth of a SD. For sentence |
| | | | Monitoring System | | children with a migration background, | | reading, the effect of aircraft noise did |
| | | | (FANOMOS), provided by | | parental involvement in school affairs, | | not reach significance, neither in the |
| | | | German Air Traffic Services | | classroom insulation, road-traffic | | unadjusted model nor in the adjusted |
| | | | | | noise, and railroad noise at school | | models |

| Spilski et al., Internoise, | Cross sectional | 439 German second- | Aircraft noise - Noise | 39 dB | Age, gender, SES, number of | Reading | A 1 dB increase in LAeq is associated |
|-----------------------------|-----------------|---------------------------|----------------------------|-------------------|-----------------------------------|-----------------|-------------------------------------------|
| 2017 | | graders from 29 schools | Exposure levels were | | children's books, non-verbal | comprehension - | with an increase in distraction of .147 |
| | | in the vicinity of | calculated on the basis of | | abilities, story comprehension, | measured using | scale points. Fully-adjusted multilevel |
| | | Frankfurt Airport, | radar data from the Flight | | phonological awareness, access to | the Suffolk | models showed that LAmax and |
| | | Germany | Track and Aircraft Noise | | phonological representations, | Reading Scale | Emergence (separately calculated |
| | | | Monitoring System | | class SES, class size, percentage | | models) are significant predictors of |
| | | | (FANOMOS) provided | | of children with a migration | | distraction of children due to aircraft |
| | | | by German Air Traffic | | background, parental involvement | | noise (mediator); which in turn has a |
| | | | Services. Road traffic and | | in school affairs, classroom | | significant impact on reading |
| | | | railway noise levels were | | insulation, road-traffic and | | performance |
| | | | estimated using a | | railway noise at school | | |
| | | | combination of | | | | |
| | | | information by local | | | | |
| | | | authorities | | | | |
| | | | | | | | |
| Seabi J et al., Expo Sci | Longitudinal | Cohort of 732 learners | Aircraft noise - To | 54.4–55.3 Leq and | Gender, deprivation, language | Reading | Results revealed no significant effect of |
| Environ Epidemiol. 2015 | field study | with a mean age of 11.1 | measure the external | 73.2–74.3 Lamax | spoken at home and groups on | comprehension - | the groups on reading comprehension |
| | | years participated at | noise surrounding the five | | reading comprehension in 2010 | measured using | across the testing periods, but |
| | | baseline measurements | schools, a SVAN 955 | | and 2011 | the Suffolk | significant effects of home language |
| | | in 2009 and 650 and | Type 1 sound lever meter | | | Reading Scale | were demonstrated on reading |
| | | 178 learners were | was utilised | | | | comprehension |
| | | reassessed after the | | | | | |
| | | relocation of the airport | | | | | |
| | | in 2010 and 2011, | | | | | |
| | | respectively | | | | | |

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| Silva et al., 2016, Applied | Cross sectional | The survey covered | A sound-level meter of | 55 dB | No adjustments made | Impact of noise - | Measurements of indoor and outdoor |
|-----------------------------|-----------------|--------------------------------------------------|-----------------------------|---------------------|----------------------------------|---------------------|------------------------------------------|
| Acoustics | | nine classes | accuracy class 1 was used | | | measured through | noise suggest that noise from the |
| | | located in three primary | to measure noise levels | | | subjective and | outside (road, schoolyard) affects the |
| | | schools | | | | objective | background noise level in classrooms |
| | | | | | | evaluation | but in varying degrees |
| | | | | | | | |
| Eagen et al., 2017 | Cross-sectional | 134 1 ¹ / ₂ hour classroom | 11 schools from 2 school | Short-term LAeq 1s, | School level assessment of free | Student distracted | Short-term exposure to aircraft noise |
| | observational | observation sessions of | districts: one school | 5s, 10s, 30s; TA | school meals, ethnicity, English | or interrupted by a | events and teacher voice masking and |
| | field study | 2nd, 3rd, 4th and 5th | district subject to noise | (time above) 55dB, | Language Learners. | noise source | voice raising behaviour but no effect on |
| | | Grade children | primarily from arrivals | 60dB and 65dB; and | | (observed). Also | student distraction |
| | | attending schools near | operations at LAX. Other | NA (number above) | | observed teaching | |
| | | Los Angeles airport | school district not heavily | 55dB, 60dB and | | voice raising and | |
| | | (LAX) | influenced by LAX | 65dB | | voice masking | |
| | | | operations | | | behaviour | |
| | | | | | | | |

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